

Assessment of Heavy Metals Pollution using Sediments and Bivalve *Brachidontes variabilis* as Bioindicator in the Gulf of Suez, Egypt

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Abstract The present work aimed to assess the quality of the coastal area of the Gulf of Suez by using sediments and bivalve *Brachidontes variabilis* to monitor heavy metal ions (Cd(II), Pb(II), Cu(II) and Zn(II)) at different stations along the western side of the gulf. The samples were collected twice per year (summer and winter) from seven stations representing different pollution sources. The concentration of studied heavy metal ions were determined using flame Atomic Absorption Spectrophotometer. The results showed that the heavy metal ions levels ranging from 1.34-2.60, 5.74-51.12, 3.37-57.91 and 14.69-95.96 µg/g (sediments) and 0.18-0.56, 0.53-2.54, 2.11-4.38 and 8.12-17.04 µg/g (*B. variabilis*) for Cd(II), Pb(II), Cu(II) and Zn(II) ions, respectively. In summer season, the high values of the most studied metals were showed, with significant difference ($p < 0.05$) for Pb only. The indices such as contamination factor (CF), pollution load index (PLI) and metal pollution index (MPI) were estimated in sediments and bivalve species to assess the degree of contamination from heavy metals at the different investigated stations. The obtained data indicated that the present study area was varied between low and moderate contamination with progressive decline in the quality of the investigated sites.

Keywords Heavy Metals; Sediments; *Brachidontes variabilis*; Gulf of Suez; Pollution indices; Bioindicator

1 Introduction

More than half the world's population lives within 60 km of the shoreline and this could rise to three-quarters by the year 2020. Adverse anthropogenic impacts on the coastal environment include eutrophication, organic, heavy metals, microbial pollution and oil spills. So, levels of contaminants in the marine environment are increasing continuously. In order to establish adequate coastal management programs, it is important to characterize the environment of concern chemically (Kesavan *et al.*, 2013). The extent of contamination can be assessed by measuring pollutant concentrations in water, sediments and aquatic organisms.

Heavy metals pollution has been a hot issue in marine environmental studies for many years. Even though metals occur naturally in the environment, due to the anthropogenic inputs which originate from various human activities the concentrations have been rising (Kanakaraju *et al.*, 2008; Lias *et al.*, 2013). Increased coastal population, rapid urbanization, oil production, industrial and tourism development, in addition to various economic activities have created numerous environmental and ecological problems in Egyptian's coastal areas. The western northern part of the Gulf of Suez, which is an industrial development area in Egypt, is suffered from different pollution problems. These lead to raise the pollutants such as heavy metals and hydrocarbons and consequently may affect the marine organisms.

Heavy metal pollution of marine biota is an environmental concern worldwide. Sessile benthic molluscs are used as quantitative biological indicators for monitoring chemical contaminants in marine environments. This monitoring tool was first proposed by Goldberg (1975) and launched as the "International Mussel Watch", promoting the use of bivalves as the main sentinel organisms (N.A.S., 1980). Molluscs (both bivalve and

gastropod) represent organisms commonly employed as bioindicators, and used as a monitor of baseline environmental metal concentrations. The use of bivalve or gastropods molluscs looks attractive as these organisms take up metals from all environmental compartments, either from the aqueous medium or through ingestion from food and inorganic particulate material and heavily concentrate them (Phillips, 1977; Bayen *et al.*, 2004; Yüzereroglu *et al.*, 2010). Moreover, they are appropriate as monitors in situ because they are sedentary or sessile, available all year long and easy to collect.

The aim of the present work describes the geographical patterns of Cd, Pb, Cu and Zn distributions in the west-northern Gulf of Suez coastal waters using *Brachidontes variabilis* as quantitative bioindicators and surface sediments pollution. It includes the investigation of seasonal effects on metal content of the *Brachidontes variabilis* biomonitors. Pollution indices using this species and sediments were used to compare and assess metal pollution (Cd, Pb, Cu and Zn) at different stations in the intertidal zones of the west-northern part of the Gulf of Suez.

2 The study area

The Gulf of Suez extends for about 250 km from the Suez port to Shadwan Island. The study area extends from the Suez Bay at the north to about 30 km down the Gulf of Suez (Fig. 1). Suez Bay is located between longitude 32° 28' and 32° 34' E and latitude 29° 54' and 29° 57' N. It is a shallow extension of the Gulf of Suez, roughly elliptic in shape with its major axis in the NE- SW direction. The average length along the major axis is about 13 Km, while the average width along the minor axis is about 8.8 Km. The mean depth is about 10 m and the horizontal surface area is 77.13 Km². The tides range varies between 80 cm at neaps and 140 cm at springs. The tidal motion is assumed to take place within all the volume of the Suez Bay due to its shallowness (Morcos, 1960; Meshal, 1970; El-Moselhy and Gabal, 2004).

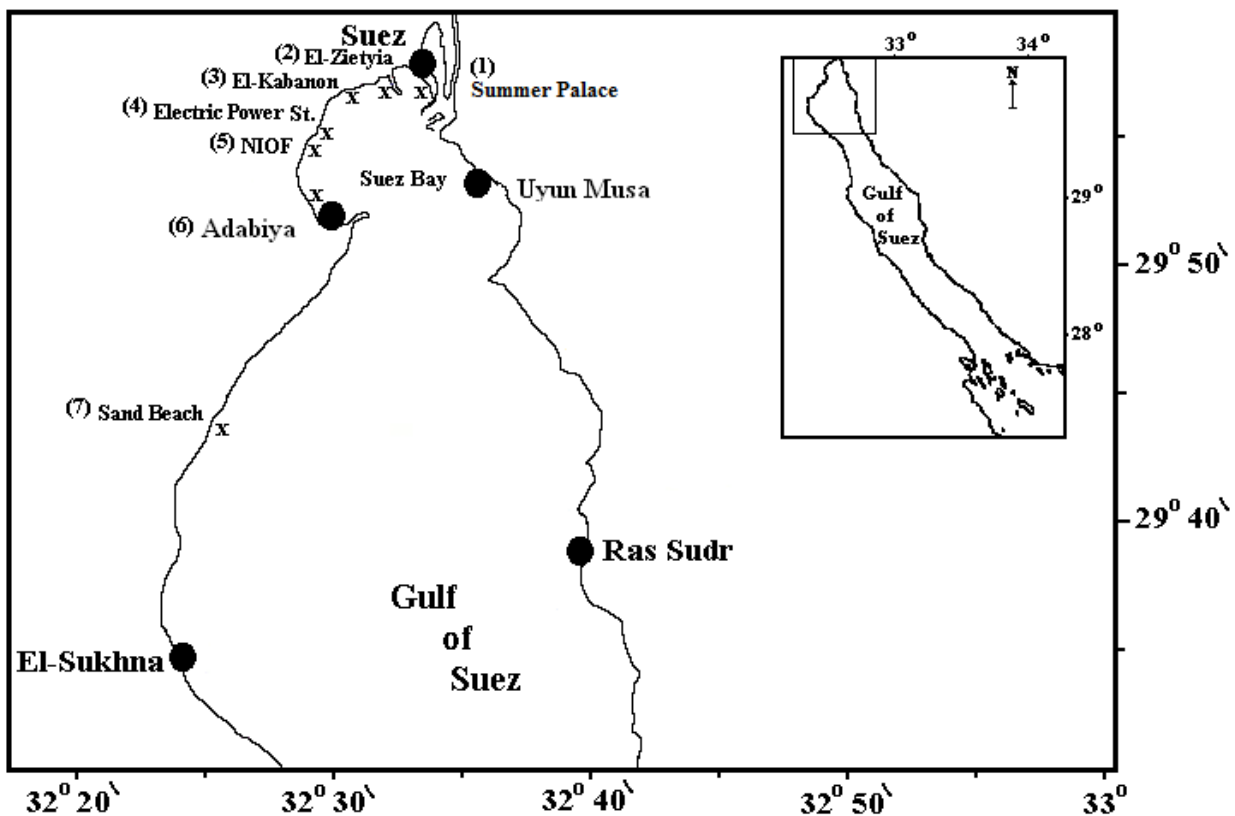


Fig. (1): Map of the northern part of the Gulf of Suez showing the sampling sites

3 Materials and Methods

3.1. Sampling stations

Seven stations have been selected in the western side of the northern part of the Gulf of Suez for sampling with regards to the sources of pollution in the present study area.

Station (1): Summer Palace Hotel beach (west of Port Tawfic Harbour): It is affected by tourist and fishing activities. It is a semi-closed area with a low wave motion and broad tide.

Station (2): El-Zeitiya Harbour: It is affected by refineries and shipping activities of petroleum tankers.

Station (3): El-Kabanon beach: This station is affected by the domestic effluents from the old sewage plant of the Suez City and effluent of thermo-power station. Its beach is characterized by a muddy substrate.

Station (4): Ataka Electric Power Plant: It receives the waste water from the power station as well as drainage water from the Fertilizers Company.

Station (5): Beach of the National Institute of Oceanography and Fisheries (NIOF): It receives water from the new sewage treatment plant of the Suez City, as well as the wastes of the Vegetable and Oil Company, Steel factory and the Trust textile factory.

Station (6): North of the Adabiya Harbour: It is affected by the shipping and trading operations in the harbour as well as the untreated sewage from a small nearby community.

Station (7): Sand Beach Resort shore: It is affected by tourist activities especially during summer season.

3.2 Sampling

Surface sandy/clay sediments and marine bivalves were collected twice during summer 2014 and winter 2015 from the intertidal area of the studying stations (Fig. 1). This area hosts a large part of the industrial and residential activities.

Surface sediment samples were collected from the intertidal area using plastic spatula. The samples were stored in plastic bags and transferred to the laboratory, air dried at room temperature and stored in plastic bags until analysis.

In the present work, one species was selected as potential quantitative bioindicators "*Brachidontes variabilis*". This species, also known as *Mytilus arabicus*, is a typical inhabitant of hard substrata and lives in shallow water (at sea level or just below) attached mostly in clusters to rocks and stones by its byssus. Its abundance seems to be negatively associated with wave exposure.

Bivalve species (*Brachidontes variabilis*) was collected from the different study stations. Organisms were collected in polyethylene bags and transferred to the laboratory in ice box, where they were identified and cleaned, then frozen at -20 °C until analysis.

3.3 Analysis

Dried Surface sediment samples were sieved and 0.5 g of the fine fraction (< 0.063 mm) was digested according to Oregioni and Aston (1984) technique; concentrated nitric, perchloric and hydrofluoric acids were added in the ratio of 3:2:1 to the sample in Teflon vessels and left overnight. The sample was then heated to 100 °C for about two hours, cooled, filtered and diluted to 25 ml with deionized water. Metals were determined by Flame Atomic Absorption Spectrophotometer (FAAS Perkin Elmer model AAnalyst 100); the results obtained were expressed in µg/g.

Brachidontes variabilis were thawed and rinsed in distilled water. The preparation of samples to determine concentration of heavy metals was carried out according to FAO (1976). Total soft tissues were separated from the shells, weight and digested at 100 °C using AR conc. nitric acid in Teflon digested vessels. Replicate samples from

each station were used for metal analysis; usually one individual of large animals represented a replicate, but composite samples from smaller size organisms were used for each replicate. Wet digested samples were filtered, diluted with deionized distilled water and analyzed for Cd, Pb Cu and Zn using Flame Atomic Absorption Spectrophotometer (FAAS Perkin Elmer model AAnalyst 100). The obtained results were expressed in $\mu\text{g/g}$ wet weight.

De-ionized water was used to prepare all solutions and blanks. All vessels and glassware were soaked in 10% nitric acid overnight and later rinsed with distilled water. Precision of the methods was verified by analysis of replicate measurements for the studied metals in the sample of sediments and marine organism. The obtained results showed precision of 7.5 – 12.3 and 4.1 – 9.8 %, respectively, for all studied metals.

3.4 Statistical Analysis

Heavy metal data in sediments and bivalve species were subjected to analysis of variance “ANOVA” to investigate the differences between stations and seasons concentration (significant values, $p \leq 0.05$ for all analysis). Duncan's multiple range test was used to further determine the position of the variance in significant results. Statistical analysis was carried out using Statistica X software packages for windows and Origin Pro 9.

4 Results and Discussion

4.1 Heavy metals in sediments

Table (1) show the values of Cd, Pb, Cu and Zn in sediment samples collected during summer and winter from different investigated stations along the western side of the northern part of the Gulf of Suez. It can be observed that, at all sampled stations, the concentrations of Zn represent the highest value in the present study area followed by Pb(II) and Cu(II) ions, while Cd(II) ion was the lowest one. As well as Cd(II), Pb(II) and Cu(II) ions showed the highest mean values during summer season; in contrast, Zn(II) was found in winter season (Fig. 2). During summer, heavy metals concentration in sediments from Gulf of Chabahar, Oman Sea were increased markedly (Bazzi, 2014); and he suggested that increasing of nutrient availability due to upwelling during summer season enhanced phytoplankton growth followed by increasing of suspended organic matter which should be involved of heavy metals enrichment of sediments. Statistically, using ANOVA analysis, only Pb(II) ion revealed significant variation between the two studied seasons ($p = 0.0426$), while the other metals showed insignificant variations. Abouhend and El-Moselhy (2015) investigated seasonal variations of heavy metals in the sediments from northern Red Sea, and recorded small temporal range of metal levels with high concentrations during winter and autumn seasons.

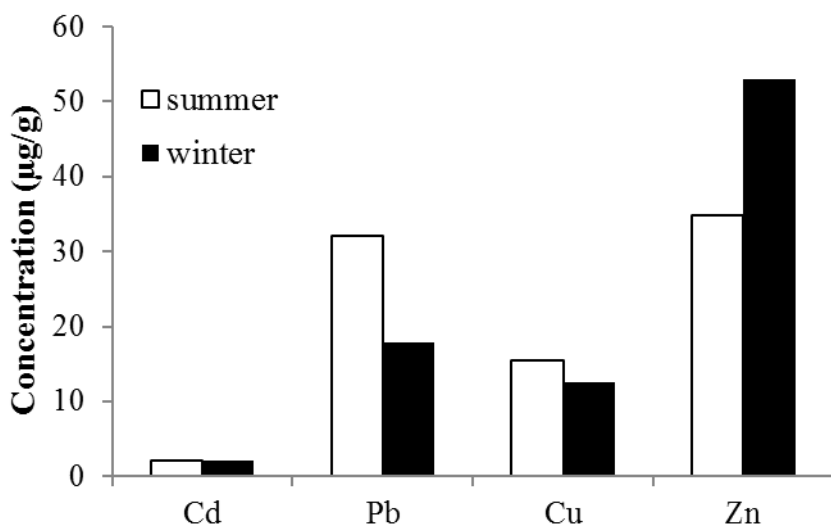


Figure (2): Mean concentrations of heavy metals ($\mu\text{g/g}$) in sediments of the northern part of the Gulf of Suez.

According to the local distribution of metals along the investigated area, it can be recorded that Pb(II), Cu(II) and Zn(II) ions exhibited their absolute highest value (51.12, 57.91 and 95.96 $\mu\text{g/g}$, respectively) at El-Zeitiya Harbour (Station 2), with annual means of 46.93, 45.16 and 75.29 $\mu\text{g/g}$, respectively. This station is exposed to pollution from oil refineries and shipping activities of petroleum tankers in addition to discharges from neighboring industries and other human activities. El-Moselhy *et al.* (1999) indicated that the land based activities and ships waiting in the area are the main sources of metal pollution in the northern part of the Gulf of Suez. In this context, El-Moselhy and Gabal (2004) found that the highest values of metals were observed at stations influenced by various pollution sources such as harbours and sewage and industrial drains. Accordingly, El-Kabanon beach (Station 3) showed maximum concentration of Cd(II) ion (2.60 $\mu\text{g/g}$ with annual mean of 2.46 $\mu\text{g/g}$). This station is influenced by the old sewage plant of the Suez City and effluent from thermo-power station in addition to current water from the northern area of the Suez Bay all over the year. In contrast, the lowest concentrations of Cd, Pb and Cu were recorded at Sand Beach Resort shore (Station 7), which is far away from any pollution sources. While, Zn(II) ion revealed its lowest value at station 3. The variations of Cd(II) and Cu(II) ions within the different studied stations were significantly different ($p = 0.0370$ and 0.0080 , respectively); while Pb(II) and Zn(II) ions showed insignificant differences ($p = 0.1438$ and 0.0775 , respectively). According to Post-Hoc Comparisons of Means "Duncan test" (Table 2), it can be noticed that the low Cd(II) ion value in station 7 was the position responsible on the significant differences ($p = 0.0063$, 0.0134 and 0.0084 with St. 3, 5 and 6, respectively). In contrast, high level of Cu(II) ion in station 2 was the liable that gave significant differences with all other stations. In addition the station 2 showed significant differences than stations 3 and 7 for Pb(II) and Zn(II) ions.

Table (1): Concentration of heavy metals and values of contamination factor (CF) in sediments collected from northern part of the Gulf of Suez during 2014-2015.

Metals	Stations	Summer	Winter	Mean \pm SD	CF
Cd	St. 1	2.07	1.95	2.01 \pm 0.08	20.10
	St. 2	2.03	1.88	1.96 \pm 0.11	19.60
	St. 3	2.60	2.31	2.46 \pm 0.21	24.60
	St. 4	2.11	2.03	2.07 \pm 0.06	20.70
	St. 5	2.51	2.12	2.32 \pm 0.28	23.20
	St. 6	2.52	2.28	2.40 \pm 0.17	24.00
	St. 7	1.34	1.84	1.59 \pm 0.35	15.90
Pb	St. 1	27.62	15.53	21.58 \pm 8.55	1.46
	St. 2	51.12	42.74	46.93 \pm 5.93	3.17
	St. 3	26.28	6.23	16.26 \pm 14.18	1.10
	St. 4	32.21	20.19	26.20 \pm 8.50	1.77
	St. 5	33.39	11.49	22.44 \pm 15.49	1.52
	St. 6	35.59	23.57	29.58 \pm 8.50	2.00
	St. 7	17.82	5.74	11.78 \pm 8.54	0.80
Cu	St. 1	6.25	7.13	6.69 \pm 0.62	0.27
	St. 2	57.91	32.40	45.16 \pm 18.04	1.81
	St. 3	5.69	4.18	4.94 \pm 1.07	0.20
	St. 4	12.57	18.12	15.35 \pm 3.92	0.61
	St. 5	12.04	6.43	9.24 \pm 3.97	0.37
	St. 6	10.62	16.09	13.36 \pm 3.87	0.53
	St. 7	3.46	3.37	3.42 \pm 0.06	0.14
Zn	St. 1	34.97	49.13	42.05 \pm 10.01	0.81
	St. 2	54.61	95.96	75.29 \pm 29.24	1.45
	St. 3	14.69	14.79	14.74 \pm 0.07	0.28
	St. 4	49.94	88.01	68.98 \pm 26.92	1.33
	St. 5	36.36	31.49	33.93 \pm 3.44	0.65
	St. 6	35.05	74.88	54.97 \pm 28.16	1.06
	St. 7	18.52	16.00	17.26 \pm 1.78	0.33

Table (2): Duncan's multiple range test showing significant differences of metals in sediments at different stations.

	St.1	St.2	St.3	St.4	St.5	St.6
	St.1					
	St.2	0.7962				
	St.3	0.0819	0.0580			
Cd	St.4	0.7782	0.6052	0.1190		
	St.5	0.1949	0.1406	0.5312	0.2707	
	St.6	0.1151	0.0819	0.7962	0.1650	0.6907
	St.7	0.0892	0.1180	0.0063*	0.0628	0.0134* 0.0084*
	St.1					
	St.2	0.0584				
	St.3	0.6270	0.0304*			
Pb	St.4	0.6830	0.0984	0.3966		
	St.5	0.9366	0.0631	0.5868	0.7301	
	St.6	0.4905	0.1414	0.2688	0.7563	0.5319
	St.7	0.3972	0.0179*	0.6819	0.2354	0.3656 0.1553
	St.1					
	St.2	0.0018*				
	St.3	0.8170	0.0015*			
Cu	St.4	0.2979	0.0048*	0.2214		
	St.5	0.7378	0.0024*	0.5880	0.4466	
	St.6	0.4083	0.0041*	0.3101	0.7932	0.5903
	St.7	0.6785	0.0013*	0.8411	0.1694	0.4731 0.2404
	St.1					
	St.2	0.1396				
	St.3	0.2126	0.0216*			
Zn	St.4	0.2117	0.7478	0.0326*		
	St.5	0.6796	0.0796	0.3597	0.1225	
	St.6	0.5155	0.3338	0.0864	0.4818	0.3182
	St.7	0.2462	0.0251*	0.8976	0.0379*	0.4063 0.1010

Heavy metals in sediments of the present study area were lie within the range of those recorded in other Egyptian coastal areas and elsewhere (El-Moselhy and Gabal, 2004; Hamed and Emara, 2006; El-Moselhy and Hamed, 2006; Sany *et al.*, 2011; Abouhend, 2013; Zaghoul, 2015 and Saad *et al.*, 2016). By comparing the present data with the background concentrations and typical levels of metals in sediments which were 0.30, 39.0, 19.0 and 98.0 $\mu\text{g/g}$ (Bryan, 1985), it can be noticed that the levels of metals in the present study showed lower concentrations, except Cd(II) ion in sediments of all stations, Pb(II) and Cu(II) ions in station 2 which were higher than the background concentration.

4.2 Heavy metals in *Brachidontes variabilis*

The concentrations of Cd, Pb, Cu and Zn in bivalve *B. variabilis* collected during summer and winter from different stations along the western side of the northern part of the Gulf of Suez are presented in Table (3). Zn was the highest recorded metal in the investigated species followed by Cu and Pb, while Cd was the lowest one. Cd, Pb and Zn exhibited their high mean values during summer season; while, Cu was found in winter season (Fig. 3).

In this context, the soft tissue of bivalve *Paphia undulata* from Lake Timsah, Suez Canal accumulated high concentrations of heavy metals during the warm months, spring-summer (El-Moselhy and Yassien, 2005). As in sediments, Pb only showed significant temporal difference ($p = 0.0198$), while other studied metals recorded insignificant variations ($p = 0.1007$, 0.1372 and 0.7968 for Cd, Cu and Zn, respectively).

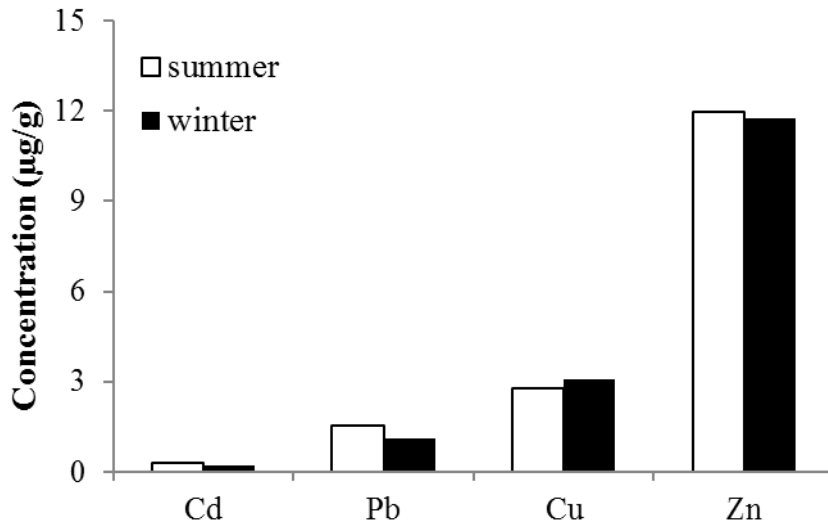


Figure (3): Mean concentrations of heavy metals ($\mu\text{g/g}$) in bivalve, *Brachidontes variabilis* collected from the northern part of the Gulf of Suez.

Temporal variations in heavy metals accumulation, which were reported in mollusca species from other regions (Szefer *et al.*, 1999; Yüzereroglu *et al.*, 2010; Singh *et al.*, 2012, 2013 and Rashida *et al.*, 2015), may be attributed to different factors such as food supply for the mollusc populations and/or runoff of particulate metal to the coastal waters of the lagoon. Seasonal fluctuations of tissue metal concentrations in molluscs may be affected by various environmental (physicochemical conditions of water) and biological factors (physiological state of organism); and have been related to a great extent to seasonal changes in flesh weight during development of gonadic tissues (Cossa and Rondeau, 1985; Joiris *et al.*, 1998; Otchere *et al.*, 2000, 2003 and Sokolowski *et al.*, 2004).

In respect to the spatial distribution of the studied metals in the bivalve *B. variabilis*, station 2 is characterized by oily sediments which suffered from pollution coming from the surrounding oil industrial area, therefore it has no any marine life appearance; accordingly, we did not found bivalve species in this station. For other stations, Cd and Zn exhibited their highest values at station 4 (0.56 and $17.04 \mu\text{g/g}$, with an annual means of 0.40 and $16.46 \mu\text{g/g}$, respectively), while Pb was found at station 5 (2.54 with an annual mean of $1.70 \mu\text{g/g}$) and Cu was in station 1 (4.38 with an annual mean of $4.06 \mu\text{g/g}$). Lowest values of the studied metals were recorded at stations 6 ($0.18 \mu\text{g Cd/g}$ and $8.12 \mu\text{g Zn/g}$), 1 ($0.53 \mu\text{g Pb/g}$) and 7 ($2.11 \mu\text{g Cu/g}$). In the present investigated bivalve, variations of Cd, Pb, Cu and Zn within the different studied stations were significantly different ($p = 0.0001$, 0.0424 , 0.0000 and 0.0000 , respectively). Post-Hoc Comparisons of Means “Duncan test” (Table 4) showed that the difference in Cd, Cu and Zn concentrations were significantly at a position of stations 1 and 4; while Pb was detected at stations 3 and 5.

The obtained results of the studied metals in soft tissues of *B. variabilis* were more or less comparable with those recorded in mollusca species by Yassien (1998), El-Moselhy *et al.* (1999), El-Moselhy and Gabal (2004), Kesavan *et al.* (2013) and Sharaf and Shehata (2015). In order to the variation in metals content in the different mollusca species may be attributed to the bioavailability of each species to uptake metals from the surrounding areas. In unpolluted and mildly polluted waters, residues of the present studied metals (Cd, Pb, Cu and Zn) are typically < 3 , 5 , 20 and $300 \mu\text{g/g}$, respectively, in base of wet and dry weight regardless of species (Chiu *et al.*, 2000; Widdows *et al.*, 2002 and Bayen *et al.*, 2004). Accordingly, the present results were lies in the range of those recorded for unpolluted and middle polluted water.

Table (3): Concentration of heavy metals and values of contamination factor (CF) in bivalve *B. variabilis* collected from northern part of the Gulf of Suez during 2014-2015.

Metals	Stations	Summer	Winter	Mean±SD	CF
Cd	St. 1	0.29±0.06	0.33±0.07	0.31±0.06	1.63
	St. 2	NF	NF	NF	NF
	St. 3	0.26±0.03	0.17±0.02	0.22±0.05	1.16
	St. 4	0.56±0.00	0.29±0.07	0.40±0.15	2.11
	St. 5	0.34±0.09	0.20±0.03	0.27±0.10	1.42
	St. 6	0.18±0.02	0.19±0.21	0.19±0.14	1.00
	St. 7	0.19±0.02	0.21±0.02	0.20±0.02	1.05
Pb	St. 1	0.53±0.35	1.43±0.17	1.02±0.53	1.15
	St. 2	NF	NF	NF	NF
	St. 3	0.82±0.40	0.97±0.14	0.89±0.30	1.00
	St. 4	2.02±0.83	0.81±0.20	1.33±0.82	1.49
	St. 5	2.54±0.47	0.85±0.28	1.70±0.97	1.91
	St. 6	1.74±0.13	1.05±0.18	1.39±0.39	1.56
	St. 7	1.59±0.62	1.47±0.20	1.54±0.46	1.73
Cu	St. 1	3.69±0.34	4.38±0.54	4.06±0.56	1.69
	St. 2	NF	NF	NF	NF
	St. 3	2.18±0.55	2.73±0.25	2.45±0.49	1.02
	St. 4	2.72±0.88	3.72±0.78	3.29±0.92	1.37
	St. 5	2.22±0.38	2.59±0.41	2.40±0.42	1.00
	St. 6	2.56±0.47	2.82±0.83	2.69±0.65	1.12
	St. 7	3.29±0.25	2.11±0.05	2.77±0.65	1.15
Zn	St. 1	12.36±1.61	15.93±1.03	14.30±2.24	1.64
	St. 2	NF	NF	NF	NF
	St. 3	9.35±1.14	9.97±0.71	9.66±0.95	1.11
	St. 4	17.04±2.08	16.02±2.21	16.46±2.04	1.89
	St. 5	13.74±1.96	10.23±0.72	11.99±2.31	1.38
	St. 6	9.29±1.34	8.12±1.60	8.70±1.52	1.00
	St. 7	9.98±1.33	10.35±0.58	10.14±1.03	1.17

NF: Not found

4.3 Assessment of heavy metals contamination

In the interpretation of the present obtained data, choice of background values plays a significant contribution. Several researchers have used the average shale values or the average crustal abundance data as reference baselines for sediment samples, these values are 0.1, 14.8, 25 and 52 µg/g for Cd, Pb, Cu and Zn, respectively (Wedepohl, 1995; Loska and Danuta, 2003 and Islam *et al.*, 2015). While for bivalve *B. variabilis*, the lowest mean values in the present study were used as baseline concentrations. The degree of contamination from heavy metals could be evaluated by determining the contamination factor (CF), pollution load index (PLI) and metal pollution index (MPI) from the following formulae 1, 2 and 3 [Tomlinson *et al.*, 1980 and Usero *et al.*, 2005].

The ratio of the measured concentration to natural abundance of a given metal had been proposed as the contamination factor (CF); thus, the CF values can monitor the enrichment of one given metal in sediments over a period of time. The obtained data for CF in sediments and bivalve species showed the range of 0.14 - 24.60 and 1.00 - 2.11, respectively, with high value for Cd (15.90 - 24.60) in sediments samples. Harikumar and Jisha (2010) stated that $CF < 1$ refers to low contamination, $1 \geq CF \geq 3$ means moderate contamination, $3 \geq CF \geq 6$

indicates considerable contamination, and $CF > 6$ indicates very high contamination. Accordingly, the present studied area was varied between low and moderate contamination regardless Cd in sediments which referred very high contamination.

Table (4): Duncan's multiple range test showing significant differences of metals in bivalve *B. variabilis* at different stations.

		St.1	St.3	St.4	St.5	St.6
Cd	St.1					
	St.3	0.0470*				
	St.4	0.0405*	0.0002*			
	St.5	0.3479	0.2386	0.0050*		
	St.6	0.0121*	0.5104	0.0000*	0.0878	
	St.7	0.0197*	0.6517	0.0001*	0.1257	0.8012
	Pb	St.1				
St.3		0.6601				
St.4		0.2790	0.1529			
St.5		0.0360*	0.0143*	0.2475		
St.6		0.2177	0.1123	0.8185	0.3248	
St.7		0.1019	0.0466*	0.4991	0.5761	0.6224
Cu		St.1				
	St.3	0.0000*				
	St.4	0.0087*	0.0083*			
	St.5	0.0000*	0.8628	0.0061*		
	St.6	0.0001*	0.4124	0.0500*	0.3519	
	St.7	0.0001*	0.3075	0.0714	0.2544	0.7883
	Zn	St.1				
St.3		0.0001*				
St.4		0.0119*	0.0000*			
St.5		0.0071*	0.0091*	0.0001*		
St.6		0.0000*	0.2518	0.0000*	0.0004*	
St.7		0.0001*	0.5617	0.0001*	0.0294*	0.1051

$$CF_{metals} = C_{metal} / C_{background} \dots\dots\dots 1$$

Where C_{metal} and $C_{background}$ are the metal concentration and the baseline metal concentration, respectively.

$$PLI = (CF_1 \times CF_2 \times CF_3 \dots\dots CF_n)^{1/n} \dots\dots\dots 2$$

Where n and CF are the metal number and contamination factor, respectively.

$$MPI = (M_1 \times M_2 \times M_3 \dots\dots M_n)^{1/n} \dots\dots\dots 3$$

Where n and M are the metal number and metal concentration, respectively.

The pollution level in different aquatic ecosystems was calculated using Pollution Load Index (PLI) (Tomlinson *et al.*, 1980; Chaudhuri *et al.*, 2007; Nomaan *et al.*, 2012; Shams El-Din *et al.*, 2014 and Ali *et al.*, 2016). It represents the number of times by which the heavy metal concentration in the samples exceeds the background concentration, and gives a cumulative indication of the overall level of heavy metal toxicity in a specific site. When the values of PLI are < 1 , it suggests a low level of pollution, a value of one indicates the presence of only baseline level of pollutants and values above one indicate progressive deterioration of the site and estuarine

quality (Tomlinson *et al.*, 1980). The PLI gave an evaluation of the overall toxicity status of the sample and also it is a consequence of the contribution of the studied four metals. The present study area showed PLI values between 0.87 and 3.57 (for sediments) with maximum at station 2, and minimum at station 7 and between 1.07 and 1.69 (for bivalve) with maximum at station 4, and minimum at station 3. The obtained PLI values indicating progressive decline in the quality of the present investigated sites (Table5).

Table (5): values of the pollution load index (PLI) and metal pollution index (MPI) in sediments and bivalve *B. variabilis* collected from the Gulf of Suez

Stations	Sediments		<i>B. variabilis</i>	
	PLI	MPI	PLI	MPI
St. 1	1.59	10.51	1.51	2.07
St. 2	3.57	23.65	---	---
St. 3	1.11	7.35	1.07	1.47
St. 4	2.34	15.48	1.69	2.32
St. 5	1.71	11.30	1.39	1.91
St. 6	2.28	15.11	1.15	1.58
St. 7	0.87	2.77	1.25	1.72

In addition to calculate the PLI values, metal pollution index (MPI) can also be used to assess the quality of the coastal areas and compare the total metal content in the different compartments of the studied area. In the present study, MPI varied from 2.77 to 23.65 (sediments) and 1.47 – 2.32 (bivalve) (Table5). MPI was previously used to evaluate the metal contamination in sediments and different marine organisms, and compare its degree between locations and within different species (Giusti *et al.*, 1999; Hamed and Emara, 2006; El-Sikaily, 2008; Abdel-Salam and Hamdi, 2014 and Ibrahim and Abu El-Regal, 2014).

According to the calculated data resulting from contamination factor and pollution indices (PLI and MPI); it can be classified the degree of contamination in the present area as following: by using sediment samples, st. 2 > st. 4 > st. 6 > st. 5 > st. 1 > st. 3 > st. 7, and by using bivalve *B. variabilis*, st. 4 > st. 1 > st. 5 > st. 7 > st. 6 > st. 3. The difference in contamination degree between sediments and bivalve species at each site may be attributed to that bivalve exposure to metals not only from the sediments or from surrounding water, but also through prey consumption which in turn bioaccumulation of heavy metals in their tissues (Wang and Fisher, 1999). Molluscs as filter feeder organisms are most frequently used to monitor the pollution of coastal water by metals (Zia and Khan, 1989). Lying in the second trophic level in the aquatic ecosystem, mollusks have long been known to accumulate trace elements in aquatic ecosystems (Phillips, 1977). However, sediments are the reservoir of metals in the aquatic systems and can be used as indicator of pollution in the coastal areas, but molluscs species “*B. variabilis* in the present study” can easily use to assess the metal pollution and is a good bioindicator for heavy metals.

5 Conclusion

The land based activities, oil industry and the different shipping activities are the main sources of metal pollution the Gulf of Suez. The level of the studied metals (Cd, Pb, Cu and Zn) in sediment samples and bivalve *B. variabilis* were varied among different stations and seasons (summer and winter). In addition to, contamination factor (CF), pollution indices (PLI and MPI) of sediments and *B. variabilis* were used to determine the degree of pollution of heavy metals at the different stations and, which indicated that the investigated area was varied between low and moderate contamination.

In conclusion, soft tissues of bivalve *Brachidontes variabilis* are suitable to be used as bioindicator for heavy metals contamination in the Gulf of Suez due to its availability to regulate and accumulate elevated concentrations of different metals and it act as a watch species in the gulf and most other Egyptian water. As well as, detailed study on the bioaccumulation level of heavy metals by other molluscs species, especially bivalve and gastropod,

from the Gulf of Suez would substantially provide complete information on their utility in monitoring program.

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